

Estrogenic chemicals in plastic and oral contraceptives disrupt development of the fetal mouse prostate and urethra

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Exposure of human fetuses to man-made estrogenic chemicals can occur through several sources. For example, fetal exposure to ethinylestradiol occurs because each year $\approx 3\%$ of women taking oral contraceptives become pregnant. Exposure to the estrogenic chemical bisphenol A occurs through food and beverages because of significant leaching from polycarbonate plastic products and the lining of cans. We fed pregnant CD-1 mice ethinylestradiol (0.1 $\mu\text{g}/\text{kg}$ per day) and bisphenol A (10 $\mu\text{g}/\text{kg}$ per day), which are doses below the range of exposure by pregnant women. In male mouse fetuses, both ethinylestradiol and bisphenol A produced an increase in the number and size of dorsolateral prostate ducts and an overall increase in prostate duct volume. Histochemical staining of sections with antibodies to proliferating cell nuclear antigen and mouse keratin 5 indicated that these increases were due to a marked increase in proliferation of basal epithelial cells located in the primary ducts. The urethra was malformed in the colliculus region and was significantly constricted where it enters the bladder, which could contribute to urine flow disorders. These effects were identical to those caused by a similar dose (0.1 $\mu\text{g}/\text{kg}$ per day) of the estrogenic drug diethylstilbestrol (DES), a known human developmental teratogen and carcinogen. In contrast, a 2,000-fold higher DES dose completely inhibited dorsolateral prostate duct formation, revealing opposite effects of high and low doses of estrogen. Acceleration in the rate of proliferation of prostate epithelium during fetal life by small amounts of estrogenic chemicals could permanently disrupt cellular control systems and predispose the prostate to disease in adulthood.

bisphenol A | ethinylestradiol | urogenital sinus

More than 60 years ago, there was speculation that exposure of male fetuses to elevated estrogen levels during fetal life could predispose men to have an enlarged prostate in old age. This hypothesis was proposed because the prostate derives from a portion of the embryonic urogenital sinus (UGS) that differentiates into the estrogen-responsive vagina in females (1). In contrast to this prediction, numerous studies have shown that high doses of diethylstilbestrol (DES) and other estrogenic chemicals inhibit prostate development in mice and rats (2). These studies were conducted because pregnant women in the 1950s and 1960s were prescribed high doses of DES based on the mistaken assumption that DES would prevent spontaneous abortion. Maternal DES administration resulted in cancer and other abnormalities of the reproductive organs in offspring, which was not detected until subsequent adulthood and after millions of human fetuses had been exposed (3–5). It is now well known that hormones can have opposite effects at low vs. high doses. Studies that include only very high doses of drugs or chemicals can miss unique effects that are observed only within a physiologically relevant low dose range (6).

We examined whether very low doses of estrogenic chemicals in drugs and consumer products (7) could affect the development of prostate ducts in male mouse fetuses. Androgen is

required for differentiation of the prostate, and our objective was to examine the modulating effect of estrogenic chemicals on the initial differentiation and growth of primary ducts in the fetal prostate. There are many possible opportunities for exposure of fetuses to estrogenic chemicals (7). An unexpected source of estrogen exposure by human fetuses is the drug ethinylestradiol, which is the estrogenic chemical used in oral contraceptives. It is estimated that each year almost 2 million of the more than 60 million women in the United States and Europe who use oral contraceptives become pregnant accidentally, primarily because of missed pills; the average is three missed pills per month and a 3% pregnancy rate per year for this population (8). Oral contraceptive pills often are taken for many months until the unplanned and unexpected pregnancy is discovered (9). Even though oral contraceptives have been used for decades, relatively little research has been conducted in experimental animals to assess effects on offspring of maternal administration of ethinylestradiol at or below the clinically relevant dose of 0.4–0.8 $\mu\text{g}/\text{kg}$ per day, based on 30 μg of ethinylestradiol per pill and body weights ranging from 37 to 75 kg (10, 11).

Bisphenol A was shown to have full activity (efficacy similar to estradiol) as an estrogenic drug in 1936 (12). Since the 1950s, bisphenol A has been used as the monomer that is polymerized to manufacture polycarbonate plastic, certain dental sealants, and the resin lining of most food and beverage cans. Bisphenol A also is used as an additive in many other products, with global capacity at >6 billion pounds per year (13). Polycarbonate is less durable than commonly believed, because the ester bond linking polymerized bisphenol A molecules can be hydrolyzed, and hydrolysis increases dramatically at high or low pH and as temperature increases. Bisphenol A thus leaches into food and beverages under the normal conditions of use of tin cans and polycarbonate plastic containers (14–16), and when polycarbonate is scratched and discolored, the rate of leaching can be very high (16). There is significant exposure of pregnant women to bisphenol A, because mean blood levels of biologically active (unconjugated) bisphenol A in human fetuses at parturition are in the range of 2–3 ng/ml (≈ 10 nM) (17), and levels in human amniotic fluid during reproductive tract differentiation in fetuses are even higher at 8 ng/ml (18).

We examined the morphology (by using computer-assisted 3D reconstruction) and cytology (by using histochemical analysis) of the fetal mouse prostate after maternal exposure for 5 days to estrogenic chemicals during the initial period of development of the primary prostate ducts; this begins on gestation day (GD) 17 in mice and occurs during the 10th week of gestation in humans.

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Abbreviations: DES, diethylstilbestrol; UGS, urogenital sinus; GD, gestation day; PCNA, proliferating cell nuclear antigen; MK5, mouse keratin 5.

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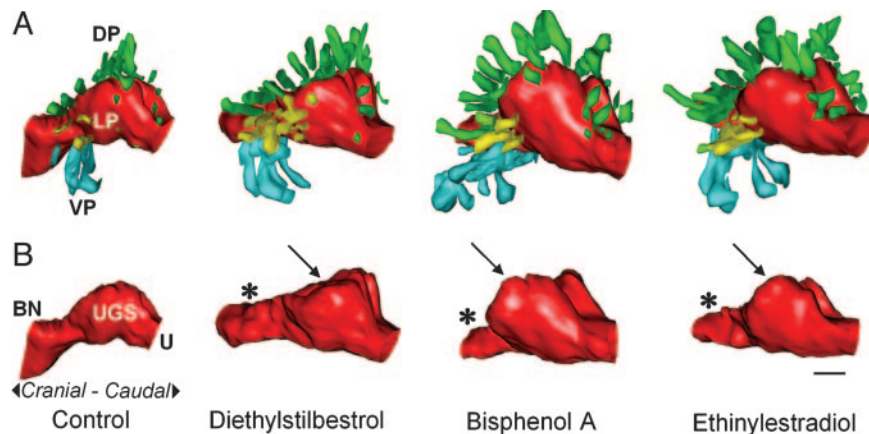


Fig. 2. 3D serial section reconstruction of the UGS from GD-19 mice exposed to low doses of estrogenic chemicals, DES, bisphenol A, and ethinylestradiol, during fetal development. The UGS depicted for each treatment was closest to the group mean for prostate duct number and size. All images are viewed from a left-lateral perspective. (A) Shown are the differences in patterns of prostate ductal development after fetal exposure to these chemicals, compared with oil-treated controls. There is a significant increase in the total number of ducts in estrogen-treated animals and a corresponding increase in overall prostate volume. (B) Shown is the marked alteration in the shape of the urethra (U, red) in the region of the bladder neck (BN), which is markedly constricted (*) in the mice exposed to the estrogenic chemicals, compared with controls. In addition, the region of the UGS (the prostatic sulcus or colliculus, arrow) associated with the development of the dorsal (DP, green) and lateral (LP, yellow) prostate ducts is enlarged, particularly by bisphenol A, compared with controls. Ventral prostate (VP, light blue). (Scale bar, 100 μm .)

ries) at a dilution of 1:200 in PBS-TX. Sections were washed in PBS and incubated in the avidin reagent as described in the ABC kit. Slides were washed again in PBS before immersion in the chromogen, diaminobenzidine, at a concentration of 0.025% made in PBS with 0.0025% hydrogen peroxide. Sections were washed in water and counterstained with hematoxylin. Slides were dehydrated, cleared, and mounted with Permount (Fisher Scientific). For MK5 staining with rabbit anti-MK5 (Covance Research Products, Berkeley, CA), antigen retrieval was performed by using procedures described by DAKO.

To determine the percentage of cells labeled for PCNA within the ducts of each region of the developing prostate, cell counts were performed on dorsal, lateral, and ventral ducts as well as the urethra. A total of 500–1,000 cells were counted in each region. Care was taken to ensure that sections selected for analysis included the entire length (both proximal and distal portions) of a duct. Sections were viewed with a BX60 microscope (Olympus, Melville, NY), and digital photographs of alternate sections throughout the entire developing gland were captured by using a DVC1301C camera (Digital Video Camera, Austin, TX). Images were analyzed by using IMAGE-PRO PLUS (Media Cybernetics, Silver Spring, MD). To determine cell counts in an area of interest, nonconvolution filters were used to enhance the outline of individual cells and differentiate between stained and unstained cells. We counted the number of cells

within each area and calculated the percent of stained vs. unstained cells within each area.

Statistical Procedures. ANOVA was conducted by using the Statistical Analysis System general linear model procedure (SAS Institute, Cary, NC). Planned comparisons of differences between controls and treatment groups were made by using the Fisher's least-squares means test in SAS if the overall analysis was statistically significant. The confidence level for rejecting the null hypothesis was $P < 0.05$. All procedures were conducted blind to ensure the absence of bias.

Results

A High Dose of DES Inhibits Prostate Morphogenesis. As predicted, administration of the high, 200- $\mu\text{g}/\text{kg}$ per day, dose of DES to pregnant mice completely inhibited the formation of ducts in the dorsal and lateral prostate in male fetuses, including the coagulating glands, which form from dorsal ducts that develop in the most cranial region of the UGS. The Müllerian ducts also were clearly evident in the DES-treated animals, suggesting that this treatment interfered with the action of Müllerian-inhibiting hormone. Relative to the oil-treated controls, this high dose of DES caused a very different pattern of outgrowths in the ventral UGS. Numerous abnormally short outgrowths were apparent throughout the entire length of the prostatic urethra, compared

Table 1. Data from reconstruction of the prostate and urethra and immunochemistry in control and estrogenic chemical-treated male fetuses

Treatment	No. of prostate ducts			Prostate duct volume, μm^2			PCNA staining, %		Cranial urethra volume, μm^3
	DL	V	DLV	DL	V	DLV	DL	V	
Control	45.4 \pm 6.3	7.2 \pm 1.3	53.0 \pm 6.7	25,616 \pm 4,773	26,531 \pm 5130	25,921 \pm 3561	36.3 \pm 3.2	35.0 \pm 7.8	35,033 \pm 8126
DES (0.1 μg)	57.2 \pm 2.5*	7.8 \pm 0.9	65.0 \pm 2.8	48,321 \pm 5,156***	56,576 \pm 2692***	51,073 \pm 3589***	64.2 \pm 2.7**	42.1 \pm 7.2	25,556 \pm 2911
BPA (10 μg)	64.2 \pm 4.6*	10.0 \pm 1.4	74.1 \pm 5.2**	50,886 \pm 6,921***	47,112 \pm 4726*	49,592 \pm 4790***	52.4 \pm 2.5*	34.1 \pm 7.5	22,767 \pm 2875*
EE (0.1 μg)	56.4 \pm 3.9*	7.6 \pm 1.2	64.0 \pm 5.1	45,508 \pm 5,215***	49,369 \pm 3947*	46,795 \pm 3688***	69.2 \pm 3.0**	40.9 \pm 9.0	21,748 \pm 2047*

All results are presented as mean \pm SEM. Shown are the number of developing epithelial ducts in the entire prostate (DLV) and for the individual dorsolateral (DL) and ventral (V) regions of the prostate on GD 19. For control, DES, and ethinylestradiol (EE) treatments, $n = 5$ fetuses per group; for bisphenol A (BPA) $n = 6$ fetuses. Prostate volume on GD 19 was calculated as the sum of all the individual cross-sectional areas based on all of the ducts in a specific region in histological sections as described in ref. 27. Urethral volume is of the initial 200 μm of the urethra beginning at the bladder neck, cranial to the region of prostate duct formation. *, $P < 0.05$; **, $P < 0.01$; ***, $P < 0.005$ vs. controls.

have opposite effects on the developing human prostate (and other tissues) relative to the high doses of DES experienced by human males whose mothers took DES during pregnancy (5). The cohort of “DES sons” may thus not be an appropriate reference population for predicting effects of exposure to low doses of estrogenic chemicals. Based on these and other findings, we propose that the use of high, pharmacological doses of estrogenic chemicals in research with experimental animals does not provide information that is relevant to understanding the role of endogenous estradiol in development (20, 32), and that the results are not relevant for predicting the effects of exposure to low levels of man-made estrogenic chemicals present in the environment (6).

An increase in prostate androgen receptors as a result of fetal exposure to very low doses of estrogenic chemicals is likely only one of many potential bases for the altered rate of epithelial proliferation. Extensive research involving experimental recombination of mesenchyme and epithelial tissues has shown that mesenchyme adjacent to specific regions of epithelium “instructs” epithelial differentiation under the control of androgen (34, 35). There have been numerous studies seeking to determine the androgen-mediated factors secreted by adjacent mesenchyme that regulate epithelial growth and differentiation. Results of these studies suggest that multiple paracrine factors, both stimulatory, such as FGF-10 (36), and inhibitory, such as TGF- β , and bone morphogenetic protein 4 (37), are involved (2, 35). A permanent derangement by estrogenic chemicals in the activity of genes that regulate cell proliferation could be a basis for the reawakening of prostate growth (both benign and malignant) during aging in men (38). There is evidence that disruption of the temporal organization of developmental events is associated with permanent functional outcomes in other organs that form ducts, such as the lungs (39), breast (40), and salivary glands (41).

Our findings show a marked increase in PCNA staining in the epithelium of the primary dorsolateral prostate ducts. Basal cells are a subset of epithelial cells found in the undifferentiated UGS that contribute to the proliferative pool of epithelial cells in the developing prostate ducts (35). Based on the known instructive mechanisms that involve mesenchymal induction of epithelial proliferation, our finding that PCNA had a similar distribution to cells that stained for MK5 suggests that estrogenic chemicals influence mesenchymal growth factors, which results in stimulation of the epithelial basal cell population. The recent identification of specific markers, such as p63 (42), which distinguishes progenitor cells in developing epithelial tissues, will permit further elucidation of the subpopulations of these cells that exhibit the greatest proliferative response to estrogenic stimulation.

In assessing whether organs in different species are homologous, functional as well as structural similarities should be assessed. With regard to functional similarities, mesenchyme from the mouse prostate has been shown to produce the appropriate regulatory factors that induce differentiation of human bladder epithelium into epithelium characteristic of the human prostate (43). The dorsolateral prostate in both human and mouse male fetuses also shows structural similarities during early fetal development, namely, the same pattern of epithelial duct formation. Based on these findings, we have proposed that the dorsolateral region of the human prostate is homologous to the dorsolateral region of the mouse prostate (2, 28). In contrast, for the ventral region of the prostate, there are differences, because unlike the mouse prostate, the human adult prostate does not contain ducts that emerge from the ventral region of the urethra. We, and others, also have reported differences between the rodent dorsolateral and ventral prostate in regulatory factors and the response to other chemicals (34, 44). The rodent dorsolateral prostate thus may be a valuable model for

predicting the effects of estrogenic chemicals on human prostate development.

The changes in the structure of the urethra induced by estrogenic chemicals may have implications for human disease. Low doses of bisphenol A, ethinylestradiol and DES caused similar morphological changes relative to oil-treated controls in the region of the paired prostatic sulci that form the lateral walls of the colliculus. This region is of particular interest with regard to estrogen-induced malformations, because the primary secretory ducts in the human prostate, which are long ducts that project into the peripheral zone, develop from these urethral ridges, and it is within these ducts that the great majority of malignant prostate tumors form in men as they age (45). The unexpected observation of a malformation of the urethra at the bladder neck (specifically, a narrowing of the urethral lumen by approximately one-third) could affect bladder function and have implications for bladder outlet obstruction disease (46).

There are now >90 published studies showing adverse effects of low doses of bisphenol A in a wide variety of experimental animals (32). For example, during postnatal life very low doses of bisphenol A have been shown to disrupt chromosomes during meiosis in mouse oocytes (47) and decrease sperm production in male rats (48). Fetal exposure to very low doses of bisphenol A accelerates postnatal growth and advances puberty (49) and also stimulates mammary gland epithelium in mice (50). A recent case-control study has shown a correlation between blood levels of bisphenol A and both obesity and polycystic ovarian disease in Japanese women (51).

With regard to ethinylestradiol, the focus of the relatively few studies of exposure by human fetuses during the critical period of reproductive organ development due to continued use of oral contraceptives during an undetected pregnancy has been on externally visible malformations at birth (reviewed in ref. 52). Our findings concerning ethinylestradiol exposure in fetal mice should be viewed from the perspective that the long-term effects of fetal DES exposure in mice and humans have been documented to be highly concordant (3–5), and we show here that the effects of the same low doses of ethinylestradiol and DES on the developing prostate and urethra are virtually identical.

Based on the general absence of grossly observable external malformations at birth, DES was considered safe for administration to millions of women during pregnancy for more than two decades. DES was subsequently found to result in serious harm to offspring that became apparent in adulthood (3–5). Similar conclusions about the safety of ethinylestradiol exposure for human fetuses based on the absence of consistent findings from studies focusing on grossly observable external malformations should be reevaluated based on several lines of evidence. Ethinylestradiol has been shown to readily pass from the maternal to fetal circulation across the primate placenta (52). Fetal exposure to DES caused uterine cancer in humans and mice (3, 4), and similar to our prostate and urethra findings here, DES and ethinylestradiol have virtually identical effects on the developing uterus in female rats (11, 53). The dose of ethinylestradiol in oral contraceptives is typically \approx 4- to 8-fold higher than the dose used in our experiment with mice that caused malformations of the urethra and altered differentiation and growth of the prostate. In summary, we propose that the data from this and other published animal studies, and the similarity to effects of low doses of DES, warrant a thorough reevaluation of the risks posed by doses of both ethinylestradiol and bisphenol A below those to which human fetuses are exposed.

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1. Zuckerman, S. (1936) *Proc. R. Soc. Med.* **29**, 1557–1567.
2. Richter, C. A., Timms, B. G. & vom Saal, F. S. (2004) in *Endocrine Disruptors: Effects on Male and Female Reproductive Systems*, ed. Naz, R. K. (CRC, Boca Raton, FL), 2nd Ed., pp. 379–410.
3. Bern, H. A. (1992) in *Chemically Induced Alterations in Sexual and Functional Development: The Wildlife/Human Connection*, Advances in Modern Environmental Toxicology, eds. Colborn, T. & Clement, C. (Princeton Scientific, Princeton), Vol. 21, pp. 9–15.
4. Newbold, R. (1995) *Environ. Health Perspect.* **103**, 83–87.
5. Swan, S. H. & vom Saal, F. S. (2001) in *Endocrine Disruptors in the Environment*, ed. Metzler, M. (Springer, Heidelberg), Vol. 3, pp. 131–170.
6. Welshons, W. V., Thayer, K. S., Taylor, J., Judy, B. & vom Saal, F. S. (2003) *Environ. Health Perspect.* **111**, 994–1006.
7. Colborn, T., vom Saal, F. S. & Soto, A. M. (1993) *Environ. Health Perspect.* **101**, 378–384.
8. Dickey, R. P. (1998) *Managing Contraceptive Pill Patients* (EMIS Publishers, Durant, OK).
9. Li, D., Daling, J. R., Mueller, B. A., Hickok, D. E., Fantel, A. G. & Weiss, N. S. (1995) *Teratology* **51**, 30–36.
10. Thayer, K. A., Ruhlen, R. L., Howdeshell, K. L., Buchanan, D., Cooke, P. S., Welshons, W. V. & vom Saal, F. S. (2001) *Hum. Reprod.* **16**, 988–996.
11. Branham, W. S., Zehr, D. R. & Sheehan, D. M. (1993) *Proc. Soc. Exp. Biol. Med.* **230**, 297–303.
12. Dodds, E. C. & Lawson, W. (1936) *Nature* **137**, 996.
13. Burridge, E. (2003) *Eur. Chem. News*, April 14–20, p. 17.
14. Brotons, J. A., Olea-Serrano, M. F., Villalobos, M., Pedraza, V. & Olea, N. (1995) *Environ. Health Perspect.* **103**, 608–612.
15. Brede, C., Fjeldal, P., Skjevraak, I. & Herikstad, H. (2003) *Food Addit. Contam.* **20**, 684–689.
16. Howdeshell, K. L., Peterman, P. H., Judy, B. M., Taylor, J. A., Orazio, C. E., Ruhlen, R. L., vom Saal, F. S. & Welshons, W. V. (2003) *Environ. Health Perspect.* **111**, 1180–1187.
17. Schonfelder, G., Wittfoht, W., Hopp, H., Talsness, C. E., Paul, M. & Chahoud, I. (2002) *Environ. Health Perspect.* **110**, A703–A707.
18. Ikezuki, Y., Tsutsumi, O., Takai, Y., Kamei, Y. & Taketani, Y. (2002) *Hum. Reprod.* **17**, 2839–2841.
19. Zalko, D., Soto, A. M., Dolo, L., Dorio, C., Rathahao, E., Debrauwer, L., Faure, R. & Cravedi, J. P. (2002) *Environ. Health Perspect.* **111**, 309–319.
20. vom Saal, F. S., Timms, B. G., Montano, M. M., Palanza, P., Thayer, K. A., Nagel, S. C., Dhar, M. D., Ganjam, V. K., Parmigiani, S. & Welshons, W. V. (1997) *Proc. Natl. Acad. Sci. USA* **94**, 2056–2061.
21. Welshons, W. V., Nagel, S. C., Thayer, K. A., Judy, B. M. & vom Saal, F. S. (1999) *Toxicol. Ind. Health* **15**, 12–25.
22. Nagel, S. C., vom Saal, F. S., Thayer, K. A., Dhar, M. G., Boechler, M. & Welshons, W. V. (1997) *Environ. Health Perspect.* **105**, 70–76.
23. Gupta, C. (2000) *Proc. Soc. Exp. Biol. Med.* **224**, 61–68.
24. vom Saal, F. S. (1989) *J. Anim. Sci.* **67**, 1824–1840.
25. Nonneman, D. J., Ganjam, V. K., Welshons, W. V. & vom Saal, F. S. (1992) *Biol. Reprod.* **47**, 723–729.
26. Timms, B. G., Petersen, S. L. & vom Saal, F. S. (1999) *J. Urol.* **161**, 1694–1701.
27. Timms, B. G., Peterson, R. E. & vom Saal, F. S. (2002) *Toxicol. Sci.* **67**, 264–274.
28. Timms, B. G., Mohs, T. J. & Didio, L. J. A. (1994) *J. Urol.* **151**, 1427–1432.
29. Iatropoulos, M. J. & Williams, G. M. (1996) *Exp. Toxicol. Pathol.* **48**, 175–181.
30. Prins, G. S. (1997) in *Prostate: Basic and Clinical Aspects*, ed. Naz, R. K. (CRC, Boca Raton, FL), pp. 247–265.
31. Wetherill, Y. B., Petra, C. E., Monk, K. R., Puga, A. & Knudsen, K. E. (2002) *Mol. Cancer Ther.* **7**, 515–524.
32. vom Saal, F. S. & Hughes, C. (April 13, 2005) *Environ. Health Perspect.*, 10.1289/ehp.7713.
33. Wozniak, A. L., Bulayeva, N. N. & Watson, C. S. (2005) *Environ. Health Perspect.* **113**, 431–439.
34. Timms, B. G., Lee, C. W., Aumuller, G. & Seitz, J. (1995) *Microsc. Res. Tech.* **30**, 319–332.
35. Marker, P. C., Donjacour, A. A., Dahiya, R. & Cunha, G. R. (2003) *Dev. Biol.* **253**, 165–174.
36. Donjacour, A. A., Thomson, A. A. & Cunha, G. R. (2003) *Dev. Biol.* **261**, 39–54.
37. Lamm, M. L. G., Podlasek, C. A., Barnett, D. H., Lee, J., Clemens, J. Q., Hebner, C. M. & Bushman, W. (2001) *Dev. Biol.* **232**, 301–314.
38. McNeal, J. E. (1978) *Invest. Urol.* **15**, 340–345.
39. Cardoso, W. V. (2000) *Dev. Dyn.* **219**, 121–130.
40. Wiesen, J. F., Young, P., Werb, Z. & Cunha, G. R. (1999) *Development (Cambridge, U.K.)* **126**, 335–344.
41. Melnick, M. & Jaskoll, T. (2000) *Crit. Rev. Oral Biol. Med.* **11**, 199–215.
42. Westfall, M. D. & Pietenpol, J. A. (2004) *Carcinogenesis* **25**, 857–864.
43. Aboseif, S., El-Sakka, A., Young, P. & Cunha, G. R. (1999) *Differentiation (Berlin)* **65**, 113–118.
44. Ko, K., Theobald, H. M., Moore, R. W. & Peterson, R. E. (2004) **79**, 360–369.
45. McNeal, J. E. (1983) *Monogr. Urol.* **4**, 1–33.
46. Streng, T., Launonen, A., Salmi, S., Saarinen, N., Talo, A., Makela, S. & Santti, R. (2001) *J. Urol.* **165**, 1305–1309.
47. Hunt, P. A., Koehler, K. E., Susiarjo, M., Hodges, C. A., Hagan, A., Voigt, R. C., Thomas, S., Thomas, B. F. & Hassold, T. J. (2003) *Curr. Biol.* **13**, 546–553.
48. Sakaue, M., Ohsako, S., Ishimura, R., Kurosawa, S., Kurohmaru, M., Hayashi, Y., Aoki, Y., Yonemoto, J. & Tohyama, C. (2001) *J. Occup. Health* **43**, 185–190.
49. Howdeshell, K. L., Hotchkiss, A. K., Thayer, K. A., Vandenberg, J. G. & vom Saal, F. S. (1999) *Nature* **401**, 763–764.
50. Markey, C. M., Luque, E. H., Munoz De Toro, M., Sonnenschein, C. & Soto, A. M. (2001) *Biol. Reprod.* **65**, 1215–1223.
51. Takeuchi, T., Tsutsumi, O., Ikezuki, Y., Takai, Y. & Taketani, Y. (2004) *Endocr. J.* **51**, 165–169.
52. Slikker, W., Bailey, J. R., Newport, G. D., Lipe, G. W. & Hill, D. E. (1982) *J. Pharmacol. Exp. Ther.* **223**, 483–489.
53. Sheehan, D. M. & Branham, W. S. (1987) *Teratog. Carcinog. Mutagen.* **7**, 411–422.